



Effect of Zn Content on the Physicochemical Characteristics and Photoactivity of ZnO Supported Activated Carbon

IS FATIMAH*, SEPTIAN P. YUDHA, KURNI DWI SOPIANA and DEVI LIA RATNASARI

Chemistry Department, Islamic University of Indonesia, Kampus Terpadu UII,
Jl. Kaliurang Km 14, Sleman, Yogyakarta, Indonesia.

*Corresponding author E-mail: isfatimah@uii.ac.id

<http://dx.doi.org/10.13005/ojc/320550>

(Received: September 22, 2016; Accepted: October 16, 2016)

ABSTRACT

In the present work, synthesis of ZnO dispersed onto activated carbon (ZnO/AC) is reported. Synthesis was conducted by using impregnation method by zinc acetate as precursor at varied Zn content. A systematic characterization of ZnO/C materials was performed using XRD, SEM, gas sorption analyzer and diffuse reflectance UV-Visible spectrophotometer. Photoactivity of the materials was studied in phenol removal via photooxidation mechanism. From the study it is confirmed that ZnO dispersion give acceleration of phenol removal refer to surface area contribution. It is no linear correlation of each physical character to the photocatalytic activity.

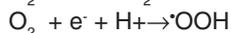
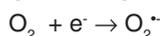
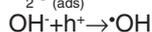
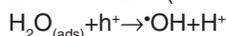
Keywords: Activated Carbon; Photocatalyst; Photooxidation; ZnO.

INTRODUCTION

Water and wastewater treatment is still the focus in environmental protection. Contamination of organic and dye compounds in aquatic system get interest refer to its potencies from several industrial activities. Several techniques were developed including adsorption, oxidation and photooxidation method. Among some methods applicable for industrial scale, in recent years photocatalytic degradation of pollutant in aquatic system became the advance technique since it can reduce the cost

the treatment due to its reusability¹⁻³. The main factor in the system is highly active photocatalyst and one of the potential photocatalysts can be elected, ZnO materials have attracted much attention due to their excellent optical properties and facile synthesis methods. ZnO itself has been reported to have activity for photodegradation and photooxidation of some organic molecules^{4,5,6,7-9}. Similar to the photocatalytic mechanism of TiO₂, the initial step in ZnO photocatalysis involves the generation of an electron-hole pair (e^-/h^+), leading to the formation of hydroxyl radicals, superoxide radical anions ($O_2^{\cdot-}$),

and hydroperoxy radicals ($\cdot\text{OOH}$) with following equations⁹



Once targeted organic molecule is attacked, radicals formed will oxidize through the mechanisms. In addition to hydroxyl radicals, superoxide radical and the positive holes are species that also possible to oxidize organic contaminants present at/or near the surface. By the mechanism it is also acceptable that as the adsorption of targeted molecule play important role for further step. With the similar value of band gap energy and chemical stability respect to TiO_2 , the activity of ZnO can be improved by attaching onto a stable and high surface area solid such as activated carbon (AC)^{3,6}. Not only by its great specific surface area and low cost, but AC also give synergistic effect on photocatalysis via the strong adsorption properties which is also the controlling factor in photocatalysis^{10,11}. From previous research, it is found that the combination of some photocatalyst with AC give synergistic effects^{2,12}. There are several methods developed and reported by previous researches. Beside of sol-gel method and the use

of organic molecule such as hydrazine as carbon source in the preparation, composite of ZnO and activated carbon was also proposed to be formed by direct infiltration. Similar approximation was also reported for the synthesis of composite from ZnO and carbon black. By the hypothetical condition of ZnO immobilizing onto AC, we report an easy method to obtain ZnO/AC composite photocatalytic material. For photocatalytic purposes, the investigation on the effect of ZnO in At the same time, the photocatalytic efficiency of ZnO/AC composite photocatalytic material was characterized by the effect of Zn content on the photocatalytic activity. The correlation of Zn content to the physicochemical character and photocatalytic activity and also some factors affecting the photocatalytic activity were studied.

MATERIALS AND METHOD

Zinc acetate dehydrate and activated carbon; the main materials for catalyst dihydrate preparation, and phenol as target molecule were provided by Merk-Millipore and were used without further purification.

The ZnO/AC was prepared as follows : zinc acetate dehydrate solution was mixed with activated carbon powder with continuous stirring for a night at

Table 1: Physicochemical character of materials

Sample	Zn content (%)	Pore Radius (Å)	Specific surface area (m ² /g)	Pore Volume (cc/g)
AC	n.d	16.79	617.684	5.187x 10 ⁻¹
ZnO/AC-2.5	2.57	17.48	556.644	4.865x10 ⁻¹
ZnO/AC-5.0	5.12	20.71	322.588	3.341x10 ⁻¹
ZnO/AC-7.5	7.80	24.24	176.631	2.141x10 ⁻¹
ZnO/AC-10.0	9.98	22.24	205.768	2.288x10 ⁻¹

Table 2: Initial rate of phenol removal over varied treatment and catalyst

Treatment	Initial rate (10 ⁻² ppm/mins)				
	AC	ZnO/AC-2.5	ZnO/AC-5.0	ZnO/AC-7.5	ZnO/AC-10.0
Photocatalysis	2.23	36.61	44.64	41.50	41.14
Adsorption	2.29	1.88	1.85	1.78	1.85
Photooxidation	19.37	41.10	77.81	48.12	67.10
Photolysis	2.84	2.84	2.84	2.84	2.84

room temperature. The concentration of zinc acetate dehydrate was set up to the theoretical Zn content of 2.5; 5.0; 7.5 and 10.0 10% wt. The sample was filtered and washed several times with distilled water until the pH values of the solution reached neutral pH, and then dried at 110 °C for 24 h in an oven. The dried sample was calcined at the temperature of 400°C and the solid obtained is designed as ZnO/AC. Refer to varied Zn content, the ZnO/SACs obtained

were labeled ZnO/C-2.5; ZnO/C-5.0, ZnO/C-7.5 and ZnO/C-10.0 respectively.

Scanning electron microscope-Energy dispersive X-ray spectroscopy (SEM-EDX JEOL) was used to investigate the morphology and chemical composition of the samples. XRD analysis was performed using X-ray diffractometer (XRD, Rigaku Bench Top-10) with Ni-filtered Cu-K α radiation. The

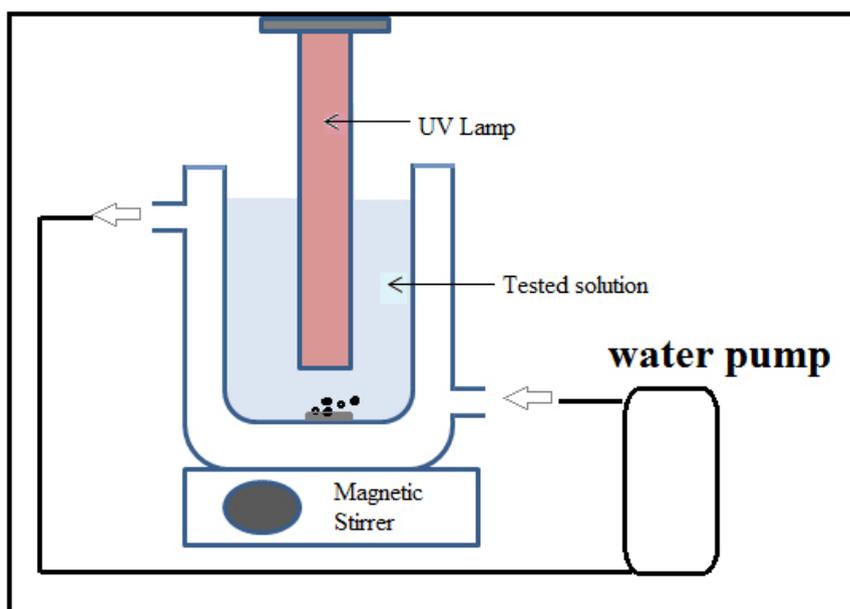


Fig. 1: Scheme of Photocatalytic reactor

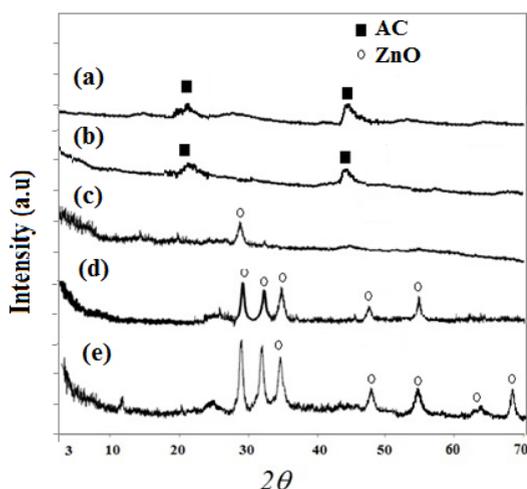


Fig. 2: XRD pattern of (a) AC (b) Zn/AC-2.5 (c) Zn/AC-5.0 (d) Zn/AC-7.5 (e) Zn/AC-10.0

specific surface area and pore structure of the ZnO/AC photocatalysts were measured by the Brunauer–Emmett–Teller (BET) method using a Micromeritics ASAP 2010 apparatus. Band gap energy of the materials was analyzed by using diffuse reflectance UV-Visible JASCO series. VG 220i-XL EscaLab was utilized for x-ray photoelectron spectroscopy (XPS) spectra identification.

Photocatalytic Study

Photocatalytic study of the photocatalysts were performed in phenol photooxidation. For photocatalytic testing a batch photoreactor with the scheme in Figure 1 was utilized. For each experiment, 0.2 g ZnO/AC photocatalyst was mixed with 500 mL of phenol solution in a batch reactor complimented with UV-lamp immersed in solution. During the test, the solution was stirred at 150 rpm.

In order to evaluate the photocatalytic activity varied methods were tested involving:

Adsorption : adsorbent addition without UV expose and without H_2O_2 as oxidant,

Photolysis : without adsorbent, addition with UV expose and without H_2O_2 as oxidant

Photocatalysis : adsorbent addition with UV expose and without H_2O_2 as oxidant, and

Photooxidation : adsorbent addition with UV expose and H_2O_2 as oxidant.

The amount of phenol (mg/L) at each time interval was calculated according to gas chromatography (GC) analysis. A Buck Scientific Gas Chromatography instrument was used.

Percentage of phenol removal is calculated based on GC responses with the following equation:

$$\text{Phenol removal(\%)} = 100\left(1 - \frac{[\text{phenol}]_{\text{solution}}}{[\text{phenol}]_{\text{initial}}}\right)$$

RESULT AND DISCUSSION

Material Preparation

In order to detect the formed ZnO on AC surface, XRD analysis was performed and the data presented in Figure 2. AC show broad and low intensity peaks at around $q = 21.8^\circ$ and 44° as reflections of the graphitic structure of AC¹³. Except for ZnO/AC-2.5 the presence of ZnO can be detected by the peaks of the ZnO nanoparticles wurtzite single phase ZnO (JCPDS Card No. 36-1451) at the 2 θ of and there is no characteristic peaks of other impurities such as $Zn(OH)_2$ detected in the diffractograms.

Based on the maximum intensity of the peak, the average diameter of the particles calculated using Scherrer formula is around 25 nm. The SEM image (Figure 3) of the ZnO/AC samples indicates some aggregates formed on surface and the higher Zn content the more aggregates formed.

The profiles data is linearly correlated with the surface profile data in Table 1 in which the

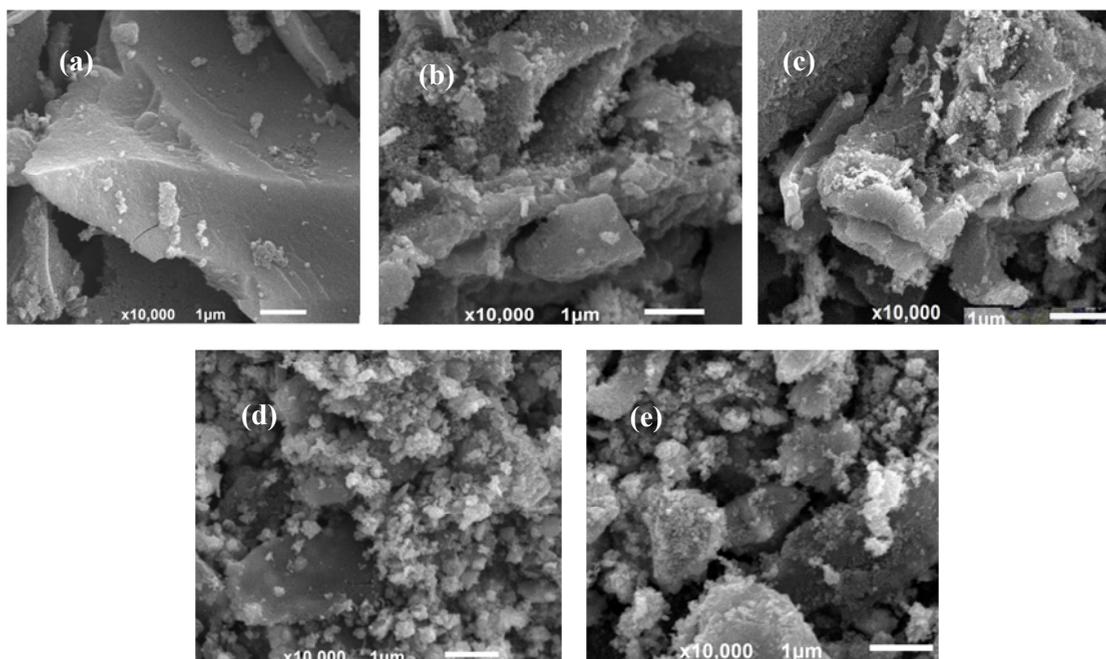


Fig. 3: SEM Profile of (a) AC (b) Zn/AC-2.5 (c) Zn/AC-5.0 (d) Zn/AC-7.5 (e) Zn/AC-10.0

calculated specific surface area, pore volume and pore radius based on N₂ adsorption-desorption data in Figure 4.

In general, increasing Zn content contributes to the decreasing specific surface area and pore volume data and getting minimum at ZnO/AC-7.5 but then those parameter for ZnO/AC-10.0 are higher than for ZnO/AC-7.5. The trend is related

to the formation of aggregates dominantly blocked the pores of activated carbon but then at the Zn percentage of 10%, ZnO aggregates create the accessible surface for adsorbate. The chemical interaction of Zn, O and C is depicted by XPS spectra of ZnO/AC-5 in Figure 5.

XPS survey spectra of the composite shows three main peaks as the proof of ZnO/AC

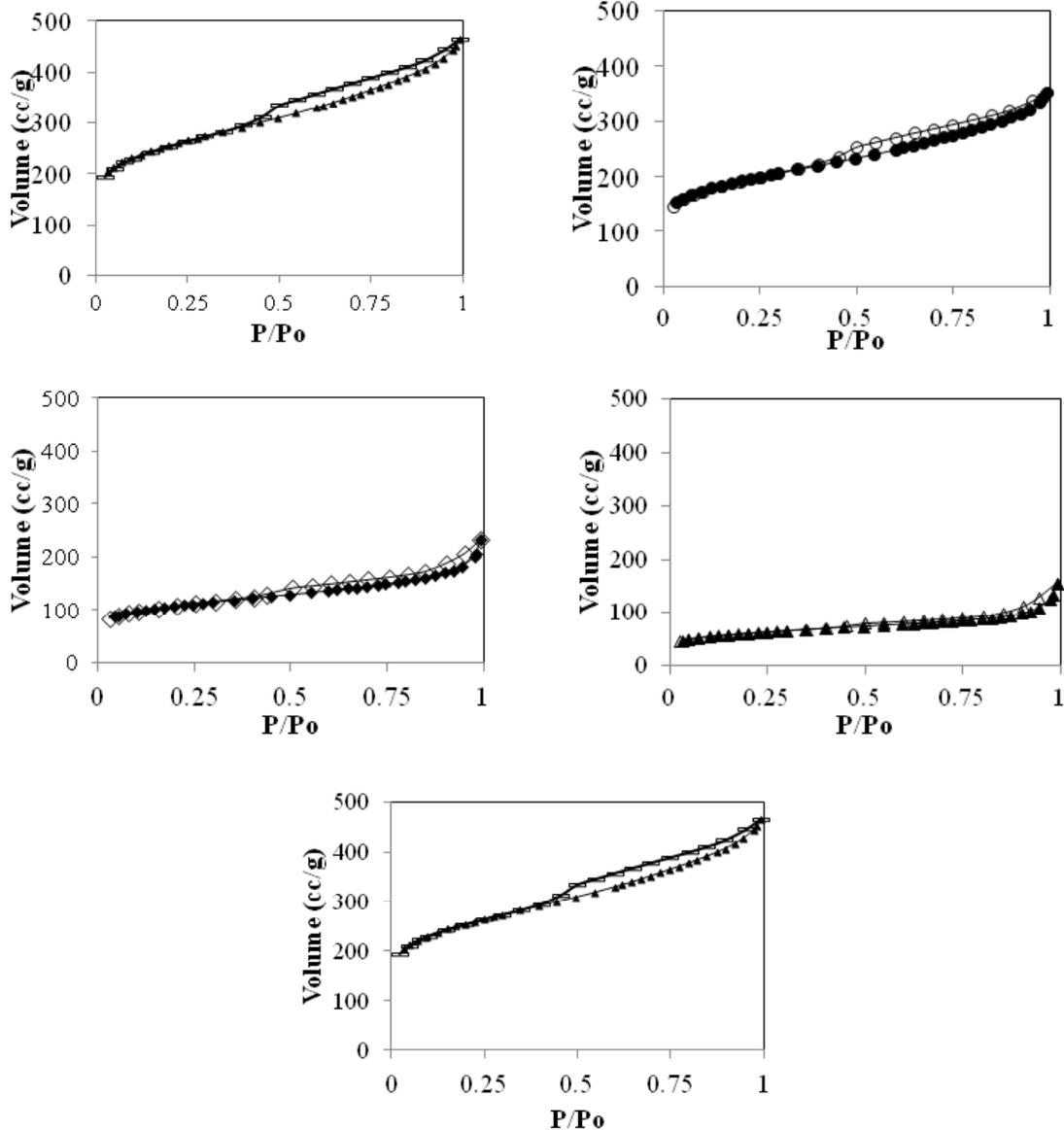


Fig. 4: Adsorption-desorption Profile of (a) AC (b) Zn/AC-2.5 (c) Zn/AC-5.0 (d) Zn/AC-7.5 (e) Zn/AC-10.0

composite, one peak at 532 eV corresponding to O 1s, one peak at 285 eV corresponding to C 1s, and two peaks at 1020–1045 eV corresponding to Zn 2p. High resolution scans of the XPS spectra of Zn 2p (1020.2 and 1045.3 eV) are correspond to Zn 2p_{3/2} and Zn 2p_{1/2} which subject to Zn–O bonding in ZnO. Furthermore the spectra at 0.500 and 0.506 keV are correspond to Zn LMM auger peak.

As different particle size formed as function of the Zn content, the effect of Zn content on the band gap energy is presented by DRUV-Vis spectra (Figure 6). From the spectra it can be concluded that the higher Zn content in the composite contribute to the red shift with the increasing intensity. This phenomenon is the quantization size effect which caused the longer wavelength due to

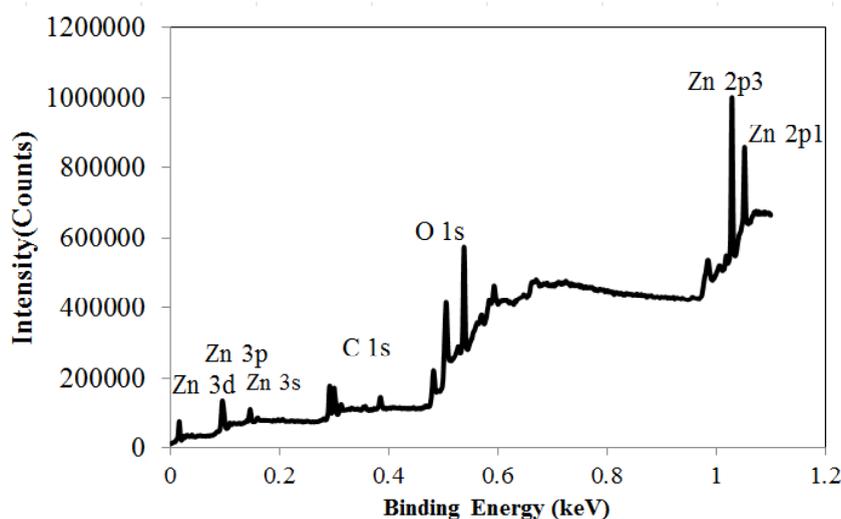


Fig. 5: XPS spectra of Zn/AC-5.0

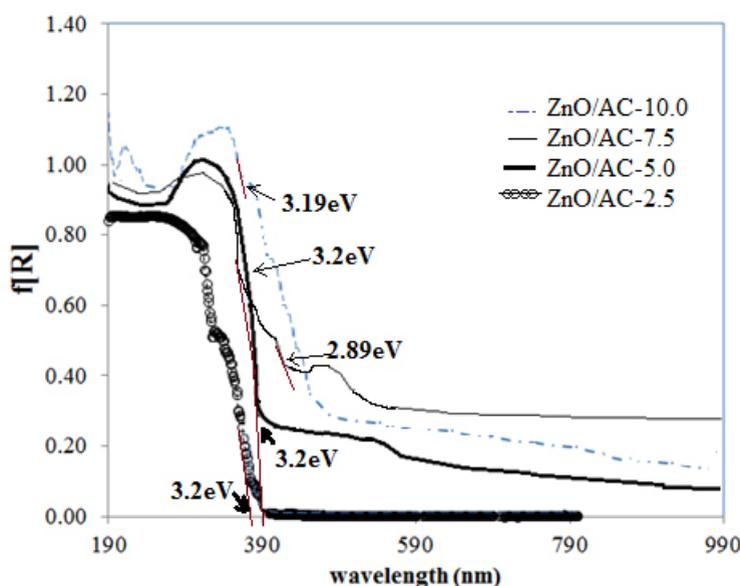


Fig. 6: DRUV-Vis spectra of materials

large particle size. Hence the light absorption and charge separation were significantly enhanced. The same trend was reported in the synthesis of ZnO nanocomposites¹⁴. As effect of the formation of different particle size of ZnO, multiple band gap values are found for ZnO/AC-7.5 and ZnO/AC-10.0 (Table 1).

Photocatalytic Activity

Photocatalytic activity of the materials was tested in varied process; adsorption, photolysis, photocatalysis, and photooxidation. Figure 7 presents the different kinetic of phenol removal over varied process over ZnO/AC-5.0 compared to AC and calculated initial rate of phenol removal by prepared materials is listed in Table 2.

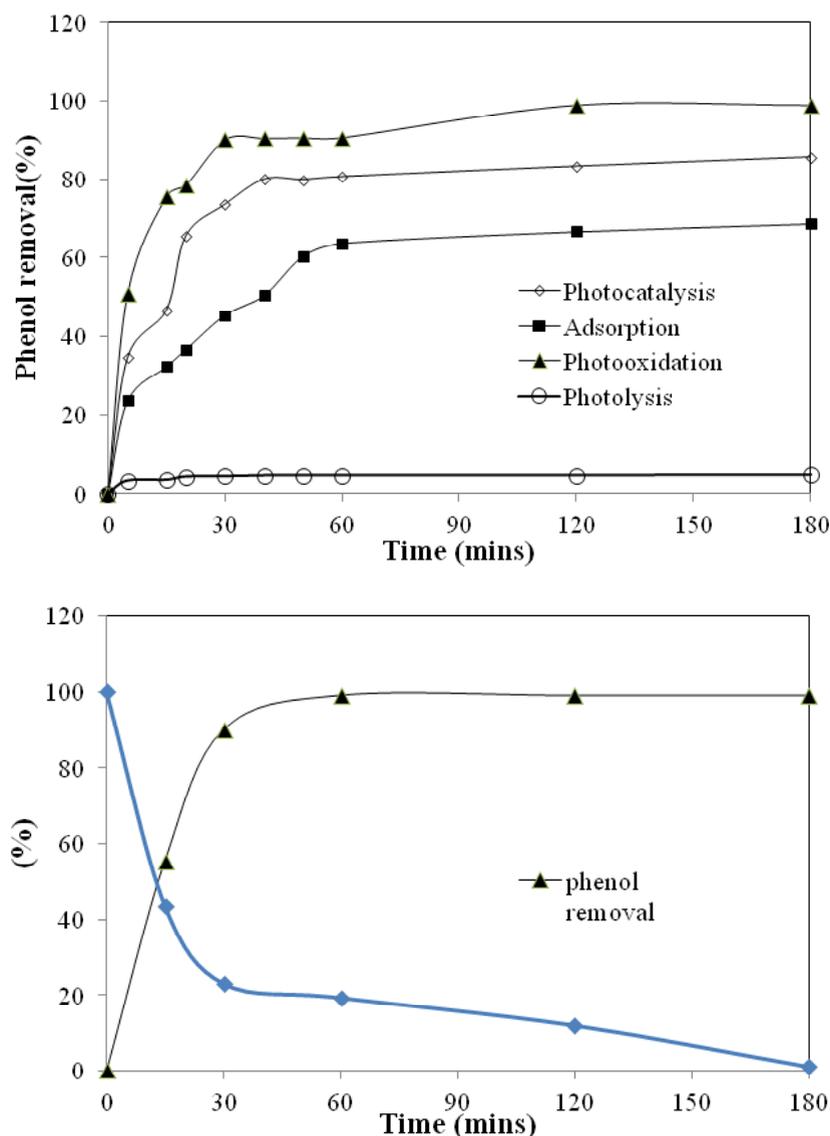
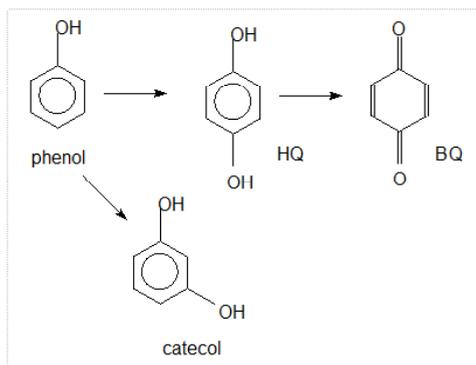


Fig. 7: (a) Kinetic of phenol removal over varied treatments (Catalyst dosage: 5 g/L; initial concentration: 10ppm, pH = 7); (b) Phenol reduction and COD reduction (Catalyst: Zn/AC-5.0; Catalyst dosage: 5 g/L; initial concentration: 10ppm, pH = 7) (Condition: Catalyst dosage: 5 g/L; initial concentration: 10ppm, pH = 7)



Degradation of phenol over photooxidation has highest rate as shown by rapid degradation percentage as increasing time of the treatment. In general the degradation rate over varied treatment is in following order: photooxidation > photocatalysis > adsorption > photolysis. From adsorption process it is noted that there is no significant difference by varying photocatalyst even there is significant difference of specific surface area as main factor in adsorption. It may related to the small amount of the adsorbent. By the additional UV expose in

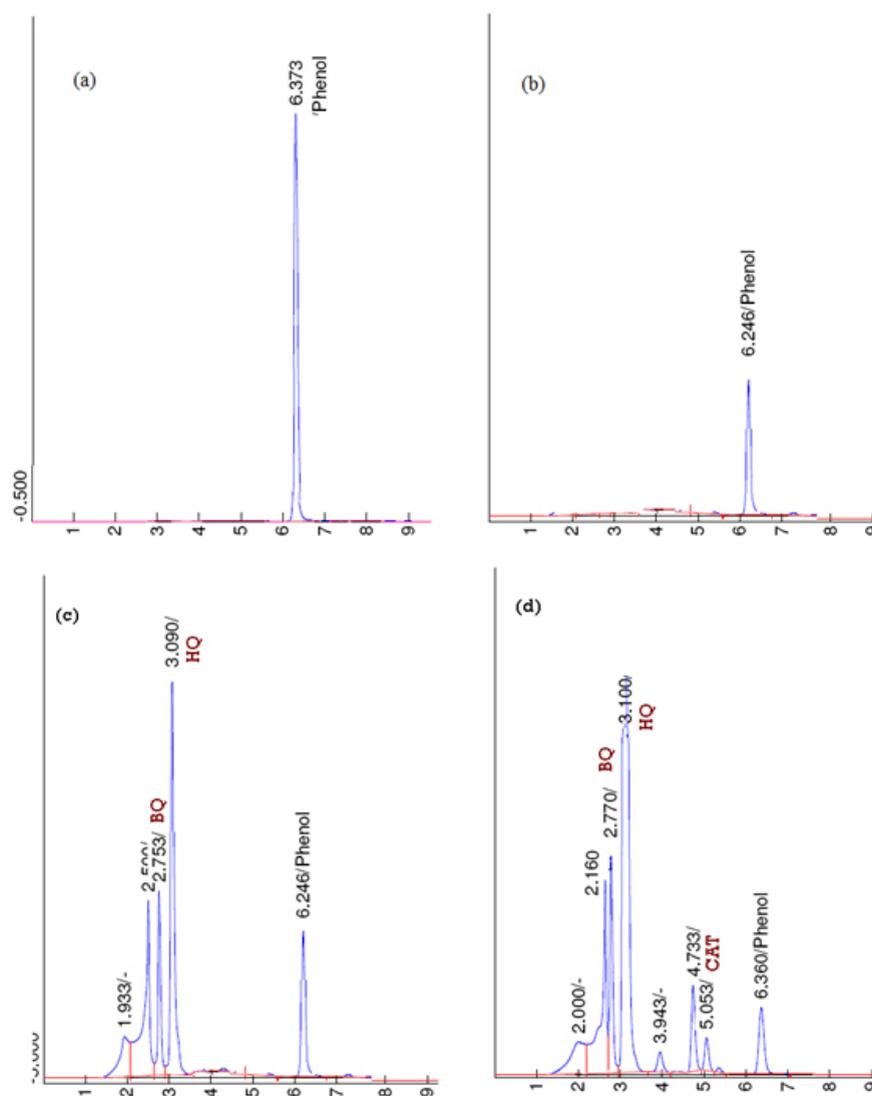


Fig. 8: Chromatogram of (a) initial solution (b) solution from adsorption treatment (c) solution from photooxidation treatment (d) solution from photocatalysis treatment (HQ: hydroquinone, CAT : catechol, BQ: benzoquinone)

photocatalysis the increasing phenol removal is found. It expresses the role of ZnO in attracting photon and producing the radicals. By comparing with phenol rate from photolysis process it is seen that phenol removal efficiency is encompassed and enhanced with the synergistic effect of both adsorption and photodegradation and with respect to the immobilization of ZnO on the porous structure of activated carbon⁶.

In advance the addition of H₂O₂ as oxidant enhance the rate. Theoretically phenol degradation in solution arises as a result of electron hole-pair formation upon irradiation of the semi-conducting oxide with appropriate light. In case recombination does not occur the electron can move to the surface of the zinc oxide as photoactive agent and combine with absorbed oxygen to form the O₂ radicals. The produced electrons from the interaction leads to radicals and super radicals formation by the interaction with oxygen and ·OH from the solvent¹⁵. It has been well established that the presence of oxidant such as H₂O₂ in the solution will accelerate the radicals formation so the photooxidation has more sufficient reactant for phenol degradation compared to photocatalysis. From the varied photocatalyst in

photocatalysis and photo-oxidation processes, it is concluded that ZnO/AC-5.0 has the best activity. The phenol removal from the solution also affects to the reduction of chemical oxygen demand(COD) as expressed in Fig.7b.

As the proof of different mechanism between photooxidation and photocatalysis is the chromatogram in Figure 8. The result of photo-oxidation and photocatalysis present the other peak as identification of products while adsorption show no other peak except phenol peak. From mass spectrometry analysis the results of photodegradation (photocatalysis and photooxidation) are identified as catechol (CAT) and hydroquinone (HQ) and benzoquinone(BQ) refer to following mechanism^{16,17}:

Effect of pH

pH of solution is an important factor influencing the interaction between target molecule and also the photocatalyst. Effect of pH on initial rate of phenol degradation is depicted in Figure 9. From the curve it is concluded base solution increases the degradation rate. Even the zero point charge of ZnO is 8.9 it is showed that at the pH lower than 8.9 for all catalyst sample the rate is not significantly

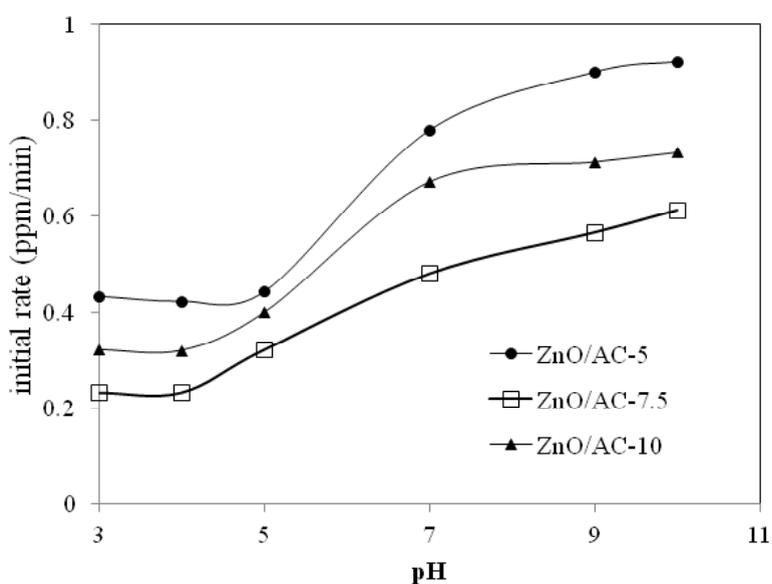


Fig. 9: Effect of pH on initial rate of phenol removal over photooxidation treatment (Condition: Catalyst dosage: 5 g/L; initial concentration: 10ppm, pH = 7)

different. The phenomenon is may related to the photocatalysis mechanism in that the presence of both $\cdot\text{OH}$ potentially accelerate the formation of radical and superadical for further oxidation step. In addition the base condition affects to catch H^+

from phenol structure. Another possible reason to be confirmed is come from the possible adsorption of OH^- onto the surface of the photocatalyst that favors to the formation of hydroxyl radicals, resulting in increasing phenol oxidation¹⁸.

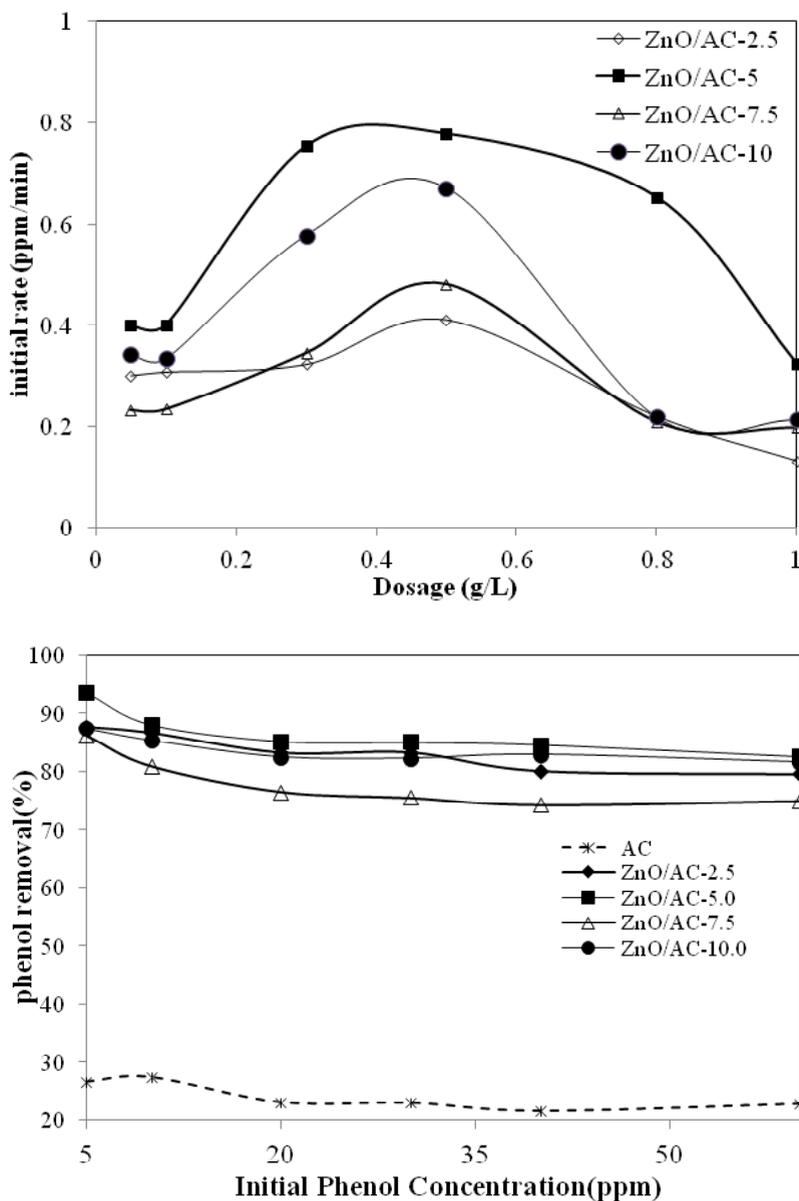


Fig.10: (a). Effect of catalyst dosage on initial rate of phenol removal over photooxidation treatment; (b) Effect of initial concentration on initial rate of phenol removal over photooxidation treatment (Condition: Catalyst dosage: 5 g/L; pH = 7)

Effect of Catalyst Dosage

Effect of catalyst dosage to the initial rate of phenol removal can be interpreted from curve in Figure 10a. Except for ZnO/AC-2.5, the photocatalysts exhibits the similar pattern of the influence of dosage to the initial rate of phenol removal. Initial rate of phenol removal is reach maximum at 0.5g/L indicating the optimum condition of photocatalysis system at the point. The possible reason for this is that the higher dosage give lower rate because of the limitation of light expose to the solution. The lower intensity of photon the lower rate of radicals formation. However in the utilization of ZnO/AC-2.5 the dosage seems to be insignificantly contribute to the increasing initial rate. The lowest content of ZnO in the photocatalyst the lower rate of radical formation obtained.

Effect of Initial Concentration

Degradation rate of phenol decreases with increasing initial concentration as shown in Figure 10b. The trend is usually found in such organic molecule photooxidation due to generally at higher amount of phenol molecules in solution the more radicals required to make oxidation complete. In relation with the data in Table 2 and the effect of pH it is also confirmed that ZnO/AC-5 give the highest activity.

By confirmation with physicochemical character of photocatalyst what can be concluded is the activity of photocatalyst is influenced by the combination of parameters factor. The highest activity was achieved by ZnO/AC-5.0 while it has similar band gap energy value with other catalysts and it also neither the highest amount of Zn nor the highest specific surface area. The photocatalyst with the highest amount of ZnO (ZnO/AC-10) does not give the highest activity probably due to its lower specific surface area among other and has the lower band gap energy values.

CONCLUSION

The study presents the preparation of ZnO/ACs expressing photocatalytic activity in phenol removal by photocatalysis mechanism. The dispersion give effect to reduce specific surface area of activated carbon and as higher Zn content, ZnO phase is identified. Photocatalytic testing in phenol removal revealed that the specific surface area, band gap energy and Zn content parameters are not individually related to the increasing activity but it is affected by the combination of those parameters. Further photocatalytic modeling is required to get the quantitative relationship.

REFERENCES

1. Bagabas, A.; Alshammari, A.; Aboud, M.F.A.; Kosslick H.; *Nanoscale.res.lett.*, **2013**, *8*, 516-525.
2. Andriantsiferana, C.; Mohamed, E.; Farouk, H. Delmas; *Environ.Technol.*, **2013**, *35*, 1-9.
3. Kyzas, G.Z.; Fu, J.; Matis, K.A.; *Materials*, **2013**, *6*,11, 312-337.
4. Sobana, N.; Thirumalai, K.; *Canadian Chem. Trans.*, **2016**, *4*, 77-89.
5. Johar, M.A.; Afzal, R.A.; Alazba, A.A.; Manzoor, U.; *Nanocomposites*, **2015**, 1-22.
6. Muthirulan, P.; Meenakshisundaram, M.; Kannan, N.; *JAdv Res.*, **2013**, *4*, 479-484.
7. Pung, S.Y.; Lee, W.P.; Aziz, A.; *Int. J. Inorg. Chem.*, **2012**, 1-9.
8. Yassýtepe, E.; Yatmaz, H.C.; Öztürk, C.; Öztürk, K.; Durana, C.; *J. Photochem. Photobiol. A*, **2018**, *198*, 1-6..
9. Daneshvar, N.; Salari, D., ; *J.Photochem. Photobiol. A*, **2004**, *162*, 317-322.
10. Forgács, E.; Cserhádi, T.; Oros, G.; *Environ Int.*, **2014**, *30*, 953-971.
11. Velasco, L.F.; Para,J.B.; Ania, C.O.; *Appl. Surf. Sci.*, **2010**, *256*, 5254-5258.
12. Asiltürk, M.; Aener, A.; *Chem.Eng.J*, **2012**, *180*, 354-363.
13. Zhao, J.; Yang, L.; Li, F.; Yu, R.; Jin, C.; *Carbon*, **2008**, *7*, 744-751.
14. Jradi, K.; Maury, C.; Daneault, C.; *Appl.Sci*, **2015**, *5*(4), 1164-1183.
15. Sathe, P.; Myint, M.T.Z.; Dobretsov, S.; Dutta, J.; *Sep.Pur.Technol.*, **2016**, *162*, 61-67.
16. Grabowska, E.; Reszczyńska, J.; Zaleska, A.; *Water Res.*, **2014**, *17*(46AD), 5453–5471.

17. Velasco, L.F.; Carmonaa, R.J.; Matos, J.; Ania, C.O.; *Carbon*. **2015**, *206*(73AD),206-2015.
18. Soltani, R.D.C.; Rezaee, A.; Khataee, A.; Safari, M.; *J. Ind. Eng.Chem.*, **2014**, *20*, 1861-1868.